Vol. 🔳 🗖

ΔΔ

MARINE ECOLOGY PROGRESS SERIES Mar Ecol Prog Ser

Snapper *Pagrus auratus* (Sparidae) home-range dynamics: acoustic tagging studies in a marine reserve

D. M. Parsons^{1,5,*}, R. C. Babcock^{1,6}, R. K. S. Hankin², T. J. Willis^{1,7}, J. P. Aitken^{3,8}, R. K. O'Dor^{3,9}, G. D. Jackson⁴

¹Leigh Marine Laboratory, University of Auckland, PO Box 349, Warkworth, New Zealand

²School of Geography and Environmental Science, University of Auckland, Private Bag 92019, Auckland 1020, New Zealand ³Department of Biology, Dalhousie University, Halifax, Nova Scotia B3H 4R2, Canada

⁴Institute of Antarctic and Southern Ocean Studies, University of Tasmania, GPO Box 252-77 Hobart, Tasmania 7001, Australia Present addresses: ⁵Department of Marine, Earth and Atmospheric Sciences, North Carolina State University, Box 8208,

Raleigh, North Carolina 27695, USA

⁶CSIRO Marine Research, Private Bag Number 5, Wembley, Western Australia 6913, Australia

⁷Scienze Ambientali, Università di Bologna, Via Tombesi dall'Ova 55, 48100 Ravenna, Italy

⁸Department of Zoology, University of Cambridge, Cambridge CB2 3EJ, United Kingdom

⁹Census of Marine Life, 1755 Massachusetts Avenue, NW #800, Washington, DC 20036, USA

ABSTRACT: The home-range size and location of reef-associated snapper *Pagrus auratus*: Sparidae were investigated by use of a radio acoustic-positioning telemetry (RAPT) system. Tags were surgically implanted in 5 snapper that were monitored every minute for a period of 5 mo, and then intermittently over a period of 1 yr. Site fidelity was high amongst these fish, with home ranges not exceeding 650 m in diameter or 139 600 m² in area. Eleven other snapper received tags by feeding and were tracked for periods of up to 2.5 d. Site fidelity was also high for these fish, with standard-ised estimates of home-range size not differing between the 2 groups. Home ranges overlapped considerably, indicating that the fish were not territorial. The location of the home range generally remained stable throughout the entire tracking period, although 1 fish relocated its home range by ~220 m. A new method of home-range estimation was developed, which matched the level of detail provided by the RAPT system, to directly estimate the time spent in an area. The relevance of this method and the residential behaviour of these fish are discussed, with reference to the general understanding of animal behaviour, previous investigations into snapper movement, and the selective capacity that may be imposed by marine reserves on fish behaviour.

KEY WORDS: *Pagrus auratus* · Snapper behaviour · Home range · Site fidelity · Residency · Utilisation distribution · New Zealand · Marine reserve

- Resale or republication not permitted without written consent of the publisher -

INTRODUCTION

The home-range parameters of animals interest biologists for 2 main reasons (Schoener 1981). First, homerange size can be related to feeding strategy, food density, resource use, metabolic demands, behaviour and efficiency of movement. Second, home-range characteristics can reflect both inter- and intraspecific interactions. Home range parameters interest conservationists and fisheries managers through their direct application to species management. For example, an understanding of fish home-range or behaviour is crucial to the effectiveness of marine reserve design (Roberts & Polunin 1991, Attwood & Bennett 1994, Holland et al. 1996, Zeller 1997, Allison et al. 1998, Woodroffe & Ginsberg 1998, Kramer & Chapman 1999, Willis et al. 125

126

127

128

184

185

186

129 2000). Whether the reserve's goal is to increase fish 130 abundance within the reserve (i.e. to protect brood 131 stock) or to supplement the adjacent fishery through 132 the emigration of fish or larval production from the 133 reserve, both of these goals could be fulfilled by a spa-134 tial restriction on fishing. However, the size of the 135 reserve relative to the mobility of the fish will influence 136 the degree to which reserve population recovery is 137 undermined by emigration to fished areas. Theoreti-138 cally, a species with intermediate dispersal capabili-139 ties, relative to reserve size, should provide a balance 140 between emigration to the fishery and accumulation of 141 brood stock (DeMartini 1993). Species with higher 142 mobility would not reside within the reserve long 143 enough to receive significant protection, while highly 144 resident species would recover the fastest but would 145 have low emigration rates to fished areas.

146 In NE New Zealand, snapper Pagrus auratus (Spari-147 dae) form the basis of the largest commercial and 148 recreational fishery (Annala et al. 1999). Snapper are 149 also the most abundant carnivorous fish within the 150 inshore areas of northern New Zealand (Paul 1976), 151 and are important at economic, cultural and ecological 152 levels. For this reason marine reserve designs in NE 153 New Zealand should optimise the effective protection 154 of snapper. A well-designed reserve would maximise 155 snapper biomass and therefore increase egg produc-156 tion (e.g. Willis et al. 2003), as well as have the poten-157 tial to benefit the fishery through emigration of adults. 158 If these goals are achieved, reserves may allow ecosys-159 tem 'recovery' by elevating snapper abundances to a 160 level where exertion of top-down processes could reg-161 ulate lower trophic levels, altering community struc-162 ture and productivity to a state reflecting the absence 163 of fishing (Babcock et al. 1999, Shears & Babcock 164 2002).

165 Despite their local importance, current knowledge of 166 home-range and space-use characteristics of snapper is 167 lacking. There is evidence suggesting that both resident 168 and mobile behaviours are exhibited by snapper. For ex-169 ample in Shark Bay, Western Australia, tagged snapper 170 from within the gulfs of Shark Bay were not recaptured 171 more than 42 km from where they were tagged, whereas 172 snapper from the open coast were recaptured up to 173 322 km from the tagging site (Moran 1987). In New 174 Zealand similar results have been gathered from tagging 175 studies. The majority of recaptures have been within 20 176 km of the tagging location, but some snapper were re-177 captured up to 418 km from the site of tagging (Paul 178 1967, Crossland 1976, Gilbert & McKenzie 1999).

Within the Cape Rodney to Okakari Point (CROP)
Marine Reserve, the site of this study, the density of
snapper above minimum legal size is 16 times greater
than in adjacent fished areas (Willis et al. 2003). As the
reserve only encompasses 5 km of coastline, the ele-

vated densities alone suggest a degree of site fidelity. Berquist (1994) investigated this residency by acoustically tagging 2 snapper within the reserve. Both fish remained within an 800 m diameter for 2 and 5 d, respectively. Using individually coded elastomer tags, Willis et al. (2001) marked 117 snapper within the CROP Reserve. Forty-nine of these fish were resignted repeatedly over several months, and the greatest distance between relocations was only 500 m.

The aim of this study was to describe the movements of 'resident' snapper within the CROP Marine Reserve, using a radio-acoustic positioning and telemetry (RAPT) system to accurately track individuals over periods of a few months. The positional fixes provided by the RAPT system were often very frequent (every minute), but not provided at regular intervals. Due to this irregular sampling frequency, we present a new method of estimating home ranges where time is used as the contouring variable. Of further interest were (1) any changes in the home-range size and location over a period of months; and (2) differences between the home range parameters of snapper that were fed tags and those that had tags surgically implanted.

MATERIALS AND METHODS

Experimental area and procedure. This study was conducted in the CROP Marine Reserve primarily from January to June 2000, although further, less frequent observations were made through to January 2001. During this time snapper were continuously tracked via the use of a RAPT system (VEMCO). This system allowed accurate positioning $(\pm 1 \text{ to } 2 \text{ m})$ (O'Dor et al. 1998) of individual fish, with a temporal resolution of minutes. Each monitored snapper contained a transmitter (pinger) that broadcast on a frequency unique to that individual. The ultrasonic signal transmitted from each fish was then received by 3 moored sono-buoys that relayed data to a land-based computer by radio signal. The computer then triangulated the position of the fish based on differences in arrival time of the signals. The sono-buoys were placed in a triangular configuration, approximately 300 m apart, within Goat Island Bay (Fig. 1). This area was chosen for its high abundance of snapper, shelter and the presence of shallow reef-habitat.

This study used V16 and V8 transmitters, also made by VEMCO. Five snapper (Table 1) received surgically implanted V16 transmitters. These V16 transmitters, ~16 mm diameter and 7.5 cm length, had a battery life conservatively estimated at 120 d (but were found to last much longer in water temperatures of 16 to 20°C). This allowed long-term detailed monitoring of snapper movements. The V8 transmitters (~8 mm diameter and 187 188 189

190

191

192

193

194

195

196

197

198

199

200

201

202

203

204

205

206

207

209

210

211

212

213

214

215

216

217

218

219 220

221

222

223

224

225

226

227

228

229

230

231

232

233

234

235

236

237

238

239

240

241

242

243

244

245

45 mm length) were small enough for snapper to swallow in situ , encased in bait, without any handling of the fish. The transmitter would be retained for ca. 2 d before passing through the body, at which time the transmitter could be relocated and retrieved using a diveroperated receiver (VUR96, VEMCO). The transmitter could then be fed to another fish. A total of 11 snapper from ca. 250 to 450 mm fork length (FL) were monitored for ca. 2 d each, by use of V8 transmitters (Table 1).

Fish capture, handling and surgery. Snapper were caught from the CROP Reserve on hook and line, using modified barbless hooks (see Willis & Millar 2001) to reduce injury and the probability of 'gut hooking'. Surgical procedures followed the methods described by Zeller (1997). After capture, each

fish was retained in an aquarium tank for 24 h to reduce stress levels before surgical insertion of ultra-sonic transmitters. Fish were anaesthetised with clove oil at 0.27 ml l⁻¹ (Munday & Wilson 1997). After the fish had become immobile it was placed in a sponge cradle and the incision area was de-scaled and then sterilised with Tamodine (Vetark products). An incision approxi-mately 2 cm long was made 1 cm from the mid-line of the fish and 2 to 3 cm anterior of the anus. The trans-mitter was then inserted into the gut cavity. The wound was sealed with nylon sutures and each fish received an injection of tetracycline antibiotic (50 mg kg⁻¹ of

Table 1. Pagrus auratus. Summary details of fish receiving V8 tags (via feeding; fish no. beginning with F) and V16 tags (implanted; fish no. beginning with S). FL: fork length

Fish no.	Fish size (mm FL)	Date released (dd/mm/yy)	Days monitored
F1	325	09/03/00	2.3
F2	400	15/03/00	2.5
F3	450	20/03/00	1.0
F4	400	20/03/00	0.8
F5	300	22/03/00	1.2
F6	300	24/03/00	1.9
F7	400	27/03/00	0.3
F8	375	28/03/00	0.2
F9	350	13/04/00	1.6
F10	400	16/05/00	2.0
F11	250	19/05/00	0.8
S4	426	24/01/00	130
S2	415	24/01/00	141
S3	532	24/01/00	141
S1	400	30/01/00	141
S5	515	04/02/00	135
S5	515	04/02/00	135





Fig. 1. Location map of North Island, New Zealand, and study area

fish). During surgery the gills were irrigated with alternate doses of pure seawater and diluted anaesthetic to ensure the fish was ventilated but remained unconscious. Each fish was then left to recover for at least 24 h in an aquarium tank before release at the site of capture. No mortality occurred during this process.

After release, manual relocations of tagged fish were made using a hand-held directional hydrophone (VR60) and a diver-operated hand-held receiver (VUR96). These were also used to record additional fish locations after the RAPT system had been removed from Goat Island Bay. All snapper were also tagged with individually coded fluorescent elastomer tags implanted in the caudal fin membranes (Willis & Babcock 1998) to allow in situ visual identification.

Data processing. Using the programming software Octave, Version 2.0 (Eaton & Rawlings 1995), the locations of each fish were recalculated from the 'R-files' generated by the RAPT system. This procedure was required because the software provided by VEMCO only recorded the average of each series of positions ('D-files'). This meant that data would have been lost through an unquantified averaging process. After all raw positions had been calculated, the data were smoothed by the following set of criteria: (1) If a location was calculated more than 1000 m from the centre of the buoy array it was deleted. VEMCO specify that the RAPT system can detect pingers up to 1 km from the buoy array (O'Dor et al. 1998), however accuracy decreases rapidly beyond this distance. (2) While the tracking system was receiving data, certain files were noted to contain obviously erroneous buoy positions, due to spurious signals during rough weather (>20 knots wind speed). Data received during these noted periods

373

374

375

376

435 436 437

438

439

440

441

442

443

444

445

446

447

448

449

450

451

452

453

454

455

456

457

458

459

460

461

462

463

464

465

466

467

468

469

470

471

472

473

474

475

476

477

478

479

480

481

482

483

484

485

486

487

488

489

490

491

492

493

494

377 were also deleted. (3) Spurious points were removed 378 by the following algorithm: Between each triplet of 379 consecutive fixes, the 2 speeds (Point 1 to Point 2 and 380 Point 2 to Point 3) were calculated. If the minimum of 381 these 2 speeds exceeded a certain maximum swim-382 ming speed, the middle point was deleted; typically, 383 Points 1 and 3 were within a metre or so of one another, 384 and Point 2 was hundreds of metres away. This process 385 was applied recursively until no 2 consecutive fixes 386 were separated by a speed exceeding the maximum 387 swimming speed. The precise value of the maximum 388 swimming speed was not critical; using values be-389 tween 1 and 10 m s⁻¹, we applied this algorithm to posi-390 tions obtained from an acoustic tag secured in a known 391 location. This resulted in only slightly differing smoothed 392 datasets. Because maximum swimming speeds for 393 snapper (or indeed other sparids) are not known, a conservative value of 4 m $\ensuremath{\mathrm{s}}^{-1}$ was used. This value is 394 395 consistent with the work of Blaxter & Dickson (1959), 396 who specified a maximum swimming speed of ~2 to 397 3 m s⁻¹ for Atlantic mackerel Scomber scombrus. In 398 addition, measurement of snapper swimming speeds 399 observed here did not exceed 0.5 m s^{-1} .

400 Home range estimation. To estimate home ranges 401 from smoothed data, the tracking area was divided into 402 a grid composed of 20×20 m bins. The amount of time 403 individual fish were detected in each of these bins was 404 then calculated using software written in MatlabTM (MathWorks 1998). This required 2 assumptions to be 405 406 made: (1) The fish swam in a straight line between con-407 secutive positional fixes as long as these fixes were not 408 more than 30 min apart. Although the RAPT system 409 attempted to locate a fish every minute, if the fish's 410 acoustic signal was obscured by sea-floor structures or 411 wave-generated noise, a fix would not be achieved. 412 Therefore, a time lapse of greater than 30 min between 413 fixes could occur. (2) The speed at which the fish swam 414 between these 2 points was constant and equal to the 415 distance divided by the time elapsed between 2 con-416 secutive positional fixes. This allowed the location of 417 the fish to be estimated between fixes as long as the 418 tracking system located the fish every 30 min or less. In 419 this way, an estimate of the amount of time a fish spent 420 within each bin of the tracking area was obtained. 421 These bin times were then contoured in ArcView, Ver-422 sion 3.2 (ESRI 1999), using the default values set for 423 proximity assignment. Each of these contours repre-424 sented the percentage of time that an individual fish 425 resided within that area. For example, the 95% con-426 tour represented the area within which a fish spent 427 95% of its time. We follow Anderson (1982) in using 428 this value to define an animal's home range. Within the 429 home range, discrete core areas were defined as areas 430 of >50% usage that were >40 m in diameter. For fish 431 that received pingers by feeding, the entire period of

tracking was represented in 1 home-range estimate. For fish that received pingers surgically, a longer timeseries of data was available. To monitor the consistency of movements, separate home-range estimates were calculated for each of these fish over 4 time periods. These were chosen in order to represent the time between new moons, as a precaution to eliminate any unknown lunar effect on snapper behaviour, and were: (1) 6 February to 6 March; (2) 6 March to 5 April; (3) 5 April to 4 May; (4) 4 May to 3 June.

RESULTS

LONG-TERM RESIDENCY

All 5 surgically tagged snapper remained attached to areas within the detection range of the tracking system (ca. 1000 m), from the time of release (January or February 2000) until the cessation of this study (June 2000). After continuous tracking ceased, 4 of these snapper were relocated 50 wk after they were originally released, using a diver-operated receiver. All relocations were within the same home ranges previously occupied by the fish. By mid-February 2001 no fish could be detected, which was probably due to the expiration of pinger batteries. By this time the pinger batteries were >200 d past their previously estimated capacity.

Home range and utilisation distribution

Surgically tagged fish

The home-range area of the 5 surgically tagged snapper varied between 13 960 and 230 000 m², whereas the area contained within the 50 % contour varied from 1700 to 14 800 m² (Table 2, Figs. 2 to 7). The largest average home range of an individual was 3.5 times greater than the smallest (i.e. 99 500 m² for Fish S1 vs 28 400 m² for Fish S5). Perhaps the best illustration of this individual variation was the contrasting movements of Fish S2 and S4. For the second monitoring period, Fish S2 spent 30.4% of its time within one 20×20 m bin, while for the third monitoring period the highest per-bin usage for Fish S4 was only 1.3%. There was no evidence of territoriality, as home ranges and core areas overlapped considerably (Figs. 2 to 7).

The size of individual home ranges changed with time, but not consistently. Between the first and last monitoring periods 3 fish increased and 2 fish decreased their home-range areas. For example, Fish S2 (Fig. 3) increased its home-range area by 24% between February and June. The size of the 50% contour did not always remain in constant proportion to the

433 434

432

497
498
499
500
501
502
502
503
504
505
506
507
508
509
510
511
510
512
513
514
515
516
517
518
519
520
520
521
522
523
524
525
526
527
528
529
520
500
531
532
533
534
534 535
534 535 536
534 535 536 537
534 535 536 537 538
534 535 536 537 538 538
534 535 536 537 538 539
534 535 536 537 538 539 540
534 535 536 537 538 539 540 541
534 535 536 537 538 539 540 541 542
534 535 536 537 538 539 540 541 542 543
534 535 536 537 538 539 540 541 542 543 544
534 535 536 537 538 539 540 541 542 543 544 545
534 535 536 537 538 539 540 541 542 543 544 545 546
534 535 536 537 538 539 540 541 542 543 544 545 546 547
534 535 536 537 538 539 540 541 542 543 544 545 546 547 548
534 535 536 537 538 539 540 541 542 543 544 545 546 547 548
 534 535 536 537 538 539 540 541 542 543 544 545 546 547 548 549
 534 535 536 537 538 539 540 541 542 543 544 545 546 547 548 549 550
 534 535 536 537 538 539 540 541 542 543 544 545 546 547 548 549 550 551
 534 535 536 537 538 539 540 541 542 543 544 545 546 547 548 549 550 551 552
 534 535 536 537 538 539 540 541 542 543 544 545 546 547 548 549 550 551 552 553

Fish no.	Monitoring period	Area within 95% contour (m²)	Area within 50% contour (m²)	Most intensive usage per 20 m bin (%)	50:95 % ratio (%)	No. of core areas	Movement of each core area (m)
S1	1	80 600	10300	6.5	12.78	1	
	2	139600	9 900	6.4	7.09	1	32.2
	3	90700	14 800	3.8	16.32	1	9.8
	4	87 100	13 600	5.4	15.61	3	9.4
	Average \pm SE	99500 ± 13500	12200 ± 1200	5.5 ± 0.63	12.95 ± 2.10	1.5 ± 0.5	17.1 ± 6.52
S2	1	43 800	2100	21.2	4.79	1	
	2	29400	1200	30.4	4.08	1	5
	3	43 600	2400	14.9	5.50	2	3
	4	35 400	1800	23.7	5.08	1	3.6
	Average \pm SE	38000 ± 3500	1900 ± 300	22.6 ± 3.2	4.87 ± 0.30	$1.3~\pm~0.25$	$3.9~{\pm}~0.51$
S3	1	54 400	7700	9	14.15	2	
	2	46 200	6000	11.9	12.99	2	1.0 and 37.0
	3	52400	5800	12.3	11.07	2	8.1 and 17.5
	4	69600	7700	12	11.06	2	12.4 and 34.0
	Average ± SE	55600 ± 5000	6800 ± 500	11.3 ± 0.77	12.32 ± 0.76	2	7.2 ± 2.88
							and 29.5 \pm 6.49
S4	1	46700	1700	5.4	3.60	1	
	2	61 200	7200	1.4	11.76	4	219.1
	3	56 200	5400	1.3	9.61	2	10.4 and 57.8
	4	60 300	5300	1.6	8.79	2	10.0 and 94.6
	Average ± SE	56100 ± 3300	4900 ± 1200	2.4 ± 1	8.45 ± 1.72	2.3 ± 0.63	79.8 ± 60.30
							and 76.2 \pm 13
S5	1	35 800	1900	4.1	5.31	2	
	2	23 000	2500	3.9	10.87	2	5.4 and 5.4
	3	24 800	2300	4.2	9.27	1	16.8
	4	30 000	2700	3.7	9.00	1	9.9
	Average \pm SE	28400 ± 2900	2300 ± 200	$4.0~\pm~0.11$	8.61 ± 1.17	$1.5~\pm~0.29$	$10.7~\pm~2.44$
Overall a	verage ± SE	55500 ± 6200	5600 ± 900	9.2 + 1.80	9.44 + 0.87	1.7 + 0.18	32.1 + 10.71

 Table 2. Pagrus auratus. Home-range summary statistics for surgically tagged snapper. Each monitoring period represents a full lunar cycle

size of the overall home range. However, the 50 % contour was always between 3.6 and 16.3 % of the size of the overall home range. Individual variation of this ratio provided a good indication of the level of residency within the home range. In general, as homerange size increased, so did the area contained within the 50 % contour. For example, Fish S1 had both the largest average value for its home range and also the largest average area within the 50 % contour (12 200 m²) (Table 2), producing an average 50:95 % ratio of 12.95 %. This was the highest 50:95 % ratio observed here, indicating that this fish used the space within its home range more evenly than the other fish tagged in this study.

548All 5 fish had more than 1 core area for at least 1 of549the monitoring periods. For all fish, except Fish S4, the550core areas were relatively stable in location, moving no551more than 37 m between monitoring periods. By visu-552ally following the shape of an individual fish's home553range over the 4 monitoring periods, it was possible to554confirm that the shape of the home range and the loca-555tion of the most intensively used areas remained rela-

tively constant (Figs. 2 to 7). Core areas appeared (Fig. 2d) and disappeared (Fig. 3c,d), but home ranges generally appeared to have been oriented around a consistent core area. This 'main' core area was not necessarily at the centre of the home range.

The exception, Fish S4, shifted its home range between 6 March and 5 April by ~220 m (Fig. 5a,b). During the second monitoring period (Fig. 5b), a series of core areas from west to east was exhibited. This presumably represented the different areas this fish resided in as it was shifting home range over the period of 1 mo. In the last 2 monitoring periods the eastern-most part of these core areas became stable. The completeness of this home-range shift is further emphasised by the fact that after April the fish did not return to its previous core area.

Fish tagged by feeding

The home range size of snapper that were fed tags varied between 3900 and $50\,329\,\,\mathrm{m^2}$, and the area



Fig. 2. *Pagrus auratus.* Home range and utilisation distributions of Fish S1 (400 mm fork length) for 4 lunar cycles between February and June 2000

within the 50 % contour varied from 122 to 2901 m² (Table 3, Fig. 7). The highest per-bin usage intensity ranged from 12.2 to 43.6 %, while the number of core areas was either 1 or 2.

The home-range sizes and areas of 50% usage for these fish were similar, but generally smaller than those of the snapper that received surgically inserted tags. Accordingly, the highest per-bin usage values were generally greater than those of the surgically tagged snapper. This was due to the short monitoring time, 0.1 to 2.3 d, relative to surgically tagged (minimum of 6.7 d) snapper. To account for these differences, 11 portions of data were selected, each with a length equal to one of the monitoring periods of fish tagged by feeding. Home ranges were then estimated for these randomly selected portions of data. Paired comparisons of these randomly selected home ranges and the home ranges of fish tagged by feeding revealed no significant difference (Wilcoxon signed rank sum, p > 0.05). This indicated that both tagging methods produced similar range estimates, but also that shorter monitoring periods underestimated the true extent of a fish's movements. The relationship between home-range size and the duration of the calculation period was Table 3. Pagrus auratus. Home-range summary statistics of snapper that received tags by feeding

Fish no.	Area within 95% contour (m²)	Area within 50% contour (m ²)	Most intensive usage per 20 m bin (%)	50:95 % ratio (%)	No. of activity centres
F1	26235	1965	16.1	7.49	1
F2	18 290	433	17.6	2.37	1
F3	17 097	1355	12.2	7.93	2
F4	12 297	813	13.3	6.61	1
F5	10 345	1138	31.2	11.00	1
F6	23 998	1342	14.6	5.59	2
F7	50 329	2901	12.7	5.76	2
F8	3877	325	40.7	8.38	1
F9	13 666	1084	17.0	7.93	1
F10	11 226	976	31.2	8.69	1
F11	11 117	122	43.6	1.10	1
Average \pm SE	18648 ± 3757	1114 ± 238	23.9 ± 3.53	6.62 ± 0.86	1.3 ± 0.14

(a)

7 807 808 809 810 811 North Reef North Reef 6th Feb. - 6th March 6th March - 5th April 812 (b) Time accounted Time accounted 813 for = 18.1 daysfor = 18.3 days814 815 816 817 818 819 820 821 822 823 824 825 Reef North Reef North 826 5th April - 4th May 4th May - 3rd June (d) 827 Time accounted Time accounted for = 11.9 days for = 12.1 days 828 829 830 831 832 833 834 835 836 837 838 839 Sono-buoys 840 0.5 Km **Contours** containing 841 percentage of usage 842 95% 843 75% 844 50% 845

Fig. 3. Pagrus auratus. Home range and utilisation distributions of Fish S2 (415 mm fork length) for 4 lunar cycles between February and June 2000

762 While using the diver-operated re-763 ceiver, numerous attempts were 764 made to visually re-sight each of the (c) 765 surgically tagged snapper. On every 766 occasion, the tagged fish would allow 767 divers to approach to a close distance, 768 as indicated by the signal strength on 769 the receiver, but would not come 770 within the diver's visual range, re-771 gardless of water clarity. These sur-772 gically tagged snapper maintained 773 this behaviour for the duration of the

further investigated. Home ranges

were estimated for randomly se-

lected portions of data with lengths

between 1 and 30 d. While there was

large variance, most probably due to

differences between individual fish,

it appeared that home range size

stabilised when ≥ 7 d of monitoring

were used in the calculation (Fig. 8).

Response to human activity

745

746

747

748

749

750

751

752

753

754

756

757

758

759

760

761

780

781

782

783

784

785

786

787

804

805

806

774 study (>5 mo). In marked contrast, 775 snapper that received tags via feed-776 ing were not as cautious, and the 777 pinger signal led to visual relocation 778 on every attempt. 779

Unaccountable time

The amount of time an individual snapper was unaccountable during a lunar tracking period varied from as little as 4.4 to as much as 22.8 d (Figs. 2 to 7). There are 4 possible

788 reasons why the RAPT system could not account for 789 snapper positions: (1) tagged snapper were moving to 790 areas outside the detection range of the system; (2) 791 tagged snapper were moving to areas where the sys-792 tem was obstructed; (3) extreme sea conditions 793 reduced the amount of time that fish could be detected; 794 and (4) the system was shut down intermittently. The 795 first of these possibilities only appeared to make a 796 major contribution to the home range estimate of Fish 797 S1 (Fig. 2). Here, part of the home range was excluded 798 from analysis by discarding locations outside its west-799 ern border. For Fish S2 and S4 (Figs. 3 & 5, respec-800 tively), a combination of explanations (2) and (3) is 801 most likely. The habitats these fish occupied were shal-802 low and complex. Therefore, when storm conditions 803 occurred, the fish were most likely to have their signals obstructed, as their habitats are areas most prone to turbulence. Indeed, the frequency of storm conditions (wave surge > 2 m) was greatest in the last 2 monitoring periods (Table 4), which could explain why these fish had the lowest percentage of time accounted for during these periods. The 3rd and 4th possibilities are likely to explain the majority of the remaining unaccounted time. When storm conditions occurred, the tracking system often produced spurious buoy and fish positions (hundreds of metres from where they should have been). Data files containing such positions were deleted to avoid incorporating errors into home-range estimates. Storm conditions also made it difficult to replace the sono-buoy batteries, resulting in the system being frequently shut down. Finally, during the last 2 monitoring periods, the system was used to con-

868

846

847

848

849

850

851

852

853

854

855

856

857

858

859

860

861

862

863

864

865

866



Fig. 4. *Pagrus auratus*. Home range and utilisation distributions of Fish S3 (532 mm fork length) for 4 lunar cycles between February and June 2000

struct a habitat map (Parsons et al. submitted). This resulted in extended lengths of time when the system was not searching for fish. When these periods of missing data were totalled (Table 5), a large proportion of these data could be accounted for, especially in the last 2 monitoring periods when bad weather and alternate use of the system were most frequent. Therefore, differences in the time accounted for were most probably related to the complications discussed above, not differences in fish behaviour. Despite the lower amount of accountable time, home-range estimates from the last 2 monitoring periods were similar, if not larger, than those produced from the earlier periods of tracking (Table 2). This suggested that when snapper could not be detected, they were still utilising space in a similar manner as when they could be detected.

DISCUSSION

Home-range size

This study presents the first estimates of snapper home range. While some previous studies have been successful in obtaining repetitive locations of individual snapper, the duration of sampling was either too short (Berguist 1994) or the number of locations too few (Willis et al. 2001) to assess snapper home-range size. The estimates of home-range size obtained here varied between 23000 and 139600 m², with corresponding maximum diameters of 190 and 620 m, respectively. These results are consistent with those of Willis et al. (2001), where residency was demonstrated over a scale of hundreds of metres in a larger sample-size of snapper (49 resighted out of 117 tagged) and a period of > 3 yr. While the logistics and cost of acoustic telemetry limited our sample size, the additional detail we provide show that the 5 fish tagged in this study were resident within the reserve for the 5 mo of monitoring. In addition, 4 of these 5 fish were located within the same individual home ranges 1 yr after release. Speculation may suggest that the reason these snapper remained resident within the CROP Reserve was due to the fish feeding activities of tourists. However, the fish tagged in this study spent either none, or a very

small, proportion of their time in areas where feeding occurred. In addition, snapper that received tags surgically

Ta	ble	4.	Wave	surge	conditions	for	each	monitoring	period
----	-----	----	------	-------	------------	-----	------	------------	--------

Monitoring period (dd/mm/yy)	Days with surge > 2 m	Accountable missing time (d)		
06/02/00-06/03/00	4	0.06		
06/03/00-05/04/00	4	2.21		
05/04/00-04/05/00	6	6.83		
04/06/00-03/06/00	9	9.54		
Number of days with accountable missing data. Explana- tions include: System being shut down due to low voltage; alternate use of the system; and deletion of data files containing spurious positions				

g

Table 5. Number of days with accountable missing data. Explanations include: (1) System being shut down due to low voltage; (3) alternate use of the system; and (2) deletion of data files containing spurious positions

Monitoring period (dd/mm/yy)	Accountable missing time (d)
06/02/00-06/03/00	0.06
06/03/00-05/04/00	2.21
05/04/00-04/05/00	6.83
04/06/00-03/06/00	9.54

would not allow divers to visually locate them. While human-derived sustenance may be important to some reserve-dwelling fish, it seems unlikely to be important here.

The home-range estimates presented here were based on the monthly monitoring periods of 5 snapper that received tags surgically. The decision to estimate home ranges over a lunar month was arbitrary; however, it did allow for observation of any changes in behaviour throughout the entire tracking period (5 mo). This decision appeared to be reasonable due to the individual consistency of home-range size (<24% change over 4 mo), and stasis of home-range location throughout the entire 5 mo tracking period (<37 m movement of core areas for all fish except S4). Snapper that received tags by feeding were not included in this estimate due to differences in the

length of the monitoring period. Because home range is a function of time as well as space, the period over which home ranges are estimated must be taken into consid-eration in order to make estimates comparable (White & Garrott 1990). Further investigation of this issue revealed that the area used by an individual snapper did not ap-pear to increase when ≥7 d of monitoring were incorpo-rated in the estimate. Therefore, home-range estimates based on periods of monitoring up to 2.3 d are not directly comparable to home ranges estimated over a month, and we recommend that at least 7 d of monitoring be used in future calculations of snapper home range. For this reason, the monitoring of snapper tagged by feeding served 2 important purposes: (1) It demonstrated that the range of movements that these fish exhibited was not dissimilar to the movements of the surgically



Fig. 5. *Pagrus auratus*. Home range and utilisation distributions of Fish S4 (426 mm fork length) for 4 lunar cycles between February and June 2000

tagged fish. This lends confidence to the idea that the surgical procedure did not drastically alter the space-use characteristics of snapper; and (2) the 11 snapper fed acoustic tags also increased the sample size of fish that expressed small-scale residency.

Utilisation distribution

The use of space within the home ranges estimated here was not uniform. Each snapper spent 50% of its time within an area that was only 3.6 to 16.3% of the total home-range size. In general, the area within which snapper spent \geq 50% of their time ranged between 1700 and 14 800 m², or 55 and 200 m in diameter. This implied that while they were observed rang-



Fig. 6. *Pagrus auratus*. Home range and utilisation distributions of Fish S5 (515 mm fork length) for 4 lunar cycles between February and June 2000

ing over an area of up to 620 m diameter, most of the time they were within an area of only 200 m diameter.
The most extreme example was Fish S4 (Fig. 5). During the first monitoring period, this fish spent 50% of its time in an area of only 1700 m², or 55 m diameter.

All surgically tagged fish had more than 1 core area in at least 1 of the monitoring periods, and these core areas were not always located at the centre of the home range. This is logical, as some areas could provide better shelter or food than others. It remains unknown whether these core areas are located where a fish resides when it is inactive (e.g. Løkkeborg et al. 2000) or whether a disproportionate amount of foraging and/or social interaction are occurring at these locations. Regardless of which resources are being utilised, they are unlikely to be distributed uniformly. Therefore, it was not unexpected that fish home ranges were irregular.

With respect to other marine fish species, only 5 studies have investigated the use of space within the home range. Four of these studies used manual tracking (Holland et al. 1993, 1996, Meyer et al. 2000, Eristhee & Oxenford 2001), while one used an automated system (Cote et al. 1998). The short duration of these studies (<62 d), the intermittent periods of tracking, and the low number of positional fixes (<1429) provided limited behavioural information below the level of home-range size estimation. When utilisation distributions were calculated in this study, time was used as the density variable, fish were continuously tracked for periods of up to 140 d, and the maximum number of fixes obtained for an individual was in excess of 475 000. Therefore, the current study presents the first accurate and long-term example of how a marine fish species occupies space on a sub-home-range level.

The behavioural variation inherent within this small sample of snapper suggested that individualised behavioural traits existed within 1 species. A further example of this variation was the daily movement of Fish S1, S2 and S4 (Figs. 2a, 4a & 6a) between their individual home ranges and 'North Reef'. These movements most commonly occurred between 10:00 and 13:00 h,

and ceased to occur altogether after March (D. M. P. unpubl. data). The characteristics of these movements were consistent with the daily and seasonal patterns that snapper exhibit while spawning (Scott et al. 1993). While it is not possible to discern the reason for these movements from this analysis, it is possible that: (1) North Reef was the site of a localised spawning aggregation within the reserve; and (2) structures such as North Reef could be used as a geographic marker for historic spawning aggregations.

Home range stability

Four of the surgically tagged fish maintained home ranges with a consistent shape and location (<37 m $\,$

1244 1245 movement between monitoring 1246 periods). Such stability was not 1247 expressed by Fish S4. Between the 1248 second and third monitoring peri-1249 ods, this fish increased the number 1250 of core areas it was using from 1 to 4 1251 (Fig. 4b). These core areas led from 1252 west to east across Goat Island Bay. 1253 Illustration of this movement using 5 1254 d portions of time (not presented) 1255 revealed that core-area shifting was 1256 a gradual process. New core areas 1257 were established by gradually 1258 increasing the use of an alternate 1259 area, while the use of the original 1260 core area was maintained. Similarly, 1261 core areas were abandoned by 1262 gradually decreasing the use of 1263 them. This fish maintained 3 core 1264 areas at one time. By the 4th moni-1265 toring period it had established, and 1266 then rejected, or was evicted from, 2 1267 core areas before settling in the 1268 eastern-most core area. This sug-1269 gested that some time between 6 1270 March and 5 April, this fish relo-1271 cated its home range by ca. 220 m. 1272 During the last monitoring period, 1273 the core area of the first monitoring 1274 period was not revisited. Therefore 1275 any resources available within the 1276 original core area were obtained 1277 from its new home range or not 1278 required at all. Kramer & Chapman 1279 (1999) speculated that relocations 1280 were most likely to occur after sev-1281 eral sampling trips from the estab-1282 lished home range. This probably 1283 was the case here.

1241

1242

1243

1284

1301

1302

Relocation events could be initi-

1285 ated by seasonal change of an environmental variable 1286 (e.g. wave exposure or the abundance of prey). At this 1287 time of year a proportion of the snapper population fol-1288 low a seasonal off-reef migration (Crossland 1976, 1289 Willis et al. 2003), which might also have some influ-1290 ence on within-reef movements. Other factors that may 1291 effect home range shifts could include the interaction 1292 with other snapper. In the current study considerable 1293 home-range overlap was observed. In addition, the 1294 high density of snapper within Goat Island Bay (Willis 1295 et al. 2003) precludes the possibility that individual 1296 home ranges of this size could be occupied exclusively. 1297 This suggests 2 things: (1) the carrying capacity of a 1298 reserve, or any other area, cannot be calculated by 1299 dividing area by the average size of a snapper home 1300

range; and (2) movements between different areas are not restricted by the possibility of entering another snapper's home range.

While Willis et al. (2001) demonstrated that snapper were resident within the CROP Reserve, results presented in the current study indicate that these fish did not leave the reserve between times of re-sighting. In short, it was possible to quantify the size and permanency of snapper home ranges. Other studies of snapper movement have also suggested that snapper were resident, but at much larger scales. With respect to the scales investigated in this study, fish movement over scales of kilometres, as described by Paul (1967) and Crossland (1982), referred to the fish as mobile, not resident. Nevertheless, from the conclusions of these

Sono-buoys 0.5 Km **Contours** containing percentage of usage 95% 75% 50%

Fig. 7. Pagrus auratus. Home range and utilisation distributions of 4 of the fish that received acoustic tags. Fork lengths: F1 = 325 mm, F2 = 400 mm, F3 = 450 mm,

F4 = 400 mm

1360

1361

1362

1364



1303



Fig. 8. Pagrus auratus. Relationship between home-range size and length of time used in the calculation. Values represent means \pm SE (n = 3)

1390 previous studies, and those presented here, it would 1391 appear that snapper are capable of exhibiting both 1392 vagile and residential behaviours. A similar pattern 1393 has been observed in the movement patterns of 1394 galjoen Coracinus capensis (Attwood & Bennett 1994). 1395 While most of the galjoen tagged were recaptured 1396 within 5 km of the release site, 17.8% were caught 1397 >25 km away, the greatest distance to recapture being 1398 1040 km.

Within the CROP Reserve, indirect evidence sug-1399 1400 gests that some snapper are wider dispersing than 1401 those tagged in this study. Willis et al. (2003) monitored 1402 the density of snapper throughout 3 NE New Zealand 1403 marine reserves and their adjacent fished areas. Con-1404 sistent seasonal fluctuations of snapper abundance, 1405 both inside and outside reserves, suggested that part 1406 of the inshore snapper population was not resident and 1407 left coastal areas sometime between April and Octo-1408 ber. The fact that similar fluctuations existed outside of 1409 reserves indicates that this pattern is probably not 1410 restricted to marine reserves.

1411 If fisheries select for different traits through in-1412 creased mortality (e.g. Hilborn & Walters 1992, Cole-1413 man et al. 1996, McGovern et al. 1998, Conover & 1414 Munch 2002, Hauser et al. 2002), then marine reserves 1415 may change this selection regime and exert their own 1416 selective pressure through decreased mortality. The 1417 observation that all snapper tagged in this study 1418 resided in areas 2 orders of magnitude smaller than 1419 previously documented (Paul 1967, Crossland 1982) 1420 may be due to the behavioural selections made by such 1421 a reserve. The explanation is as follows: Within the 1422 snapper population a continuum of mobility behaviour 1423 exists. Within reserves, the fish with the highest ten-1424

dency to exhibit residential behaviour are favoured. This is due to the small size of established reserves (<9 km²) and the heavy fishing pressure on their boundaries (T. J. W. pers. obs.). Any snapper of higher mobility would therefore spend at least some time outside of the reserve, increasing the chance of capture. If all snapper were uniformly as mobile as described by Paul (1967) and Crossland (1982), then it is likely that snapper abundances would not have responded as positively to protection within reserves of the current size (Willis et al. 2003). Those estimates reflect the average mobility of a population whose behavioural distribution may have been altered by exploitation, whereas the estimates presented in this study represent individual estimates from a population with behavioural traits that may have been affected by a lack of

exploitation. This scenario illustrates 2 important points: (1) within a species, assumptions about homogeneous behaviour cannot always be made (Willis et al. 2001), and management decisions, rather than being based on such assumptions, are likely to have unexpected and possibly unfavourable consequences; and (2) a marine reserve's potential to replenish adjacent fisheries will be dependent on the reproductive and growth potential of the individuals it selects for.

Acknowledgements. We thank the Department of Conservation and the Vice Chancellors Committee of the University of Auckland for providing funding and financial assistance. A loan of equipment was gratefully received from G. Russ of James Cook University (VUR 96). We would also like to thank G. Allen for developing the home range model, D. Webber for assistance with position calculation, N. Tolimieri and 3 reviewers for comments on the manuscript, and D. Feary and numerous other voluntary field assistants.

LITERATURE CITED

- Allison GW, Lubchenco J, Carr MH (1998) Marine reserves are necessary but not sufficient for marine conservation. Ecol Appl 8(Suppl):79–92
- Anderson DJ (1982) The home range: a new nonparametric estimation technique. Ecology 63:103–112
- Annala JH, Sullivan KJ, O'Brien CJ (1999) Report from the Fishery Assessment Plenary, April 1999: stock assessments and yield estimates (unpublished report held in NIWA library, Wellington)
- Attwood CG, Bennett BA (1994) Variation in dispersal of galjoen (*Coracinus capensis*) (Teleostei: Coracinidae) from a marine reserve. Can J Fish Aquat Sci 51:1247–1257
- Babcock RC, Kelly S, Shears NT, Walker JW, Willis TJ (1999) Changes in community structure in temperate marine reserves. Mar Ecol Prog Ser 189:125–134
- Berquist RM (1994) Patterns of activity and movement in New

1432

1433

1434

1435

1436

1437

1438

1439

1440

1441

1442

1443

1444

1445

1446

1447

1448

1449

1450

1451

1452

1453

1454

1455

1456

1457

1458

1459

1460

1461

1462

1463

1464

1465

1466

1467

1468

1469

1470

1471

1472

1473

1474

1475

1476

1477

1478

1479

1480

1481

1482

1483

1484

1485

1486

1425

1365

1366 1367

1368

1369

1370

1371

1372

1373

1374

1375

1376

1377

1378

1379

1380

1381

1382

1383

1384

1385

1386

1387

1388

1389

12

zealand snapper, <i>Pagrus auratus</i> . MSc thesis, University of Auckland	1
Blaxter JHS, Dickson W (1959) Observations of swimming	
speeds of fish. J Cons Int Explor Mer 24:472–479	
Burt WH (1943) Territoriality and home range concepts as	
applied to mammals. J Mamm 24:346–352	ľ
Cole RG, Ayling TM, Creese RG (1990) Effects of marine	
reserve protection at Goat Island, northern New Zealand.	
Coleman EC Koonig CC Huntsman CR Musick IA and 5	
others (2000) Long-lived reef fishes: the grouper-snapper	ז
complex. Fisheries 25:14–21	-
Conover DO, Munch SB (2002) Sustaining fisheries yields	1
over evolutionary time scales. Science 297:94–96	
Cote D, Scruton DA, Niezgoda GH, McKinley RS, Rowsell DF,	
Lindstrom RT, Ollerhead LMN, Whitt CJ (1998) A coded	(
acoustic telemetry system for high precision monitoring of	
nearshore nursery habitat of juvenile Atlantic cod (Gadus	
Morhua). Mar Technol Soc J 32:54–62	ł
Crossland J (1981) The biology of the New Zealand snapper.	
NZ Fish Res Div Occl Pub No 23	
Crossland J (1982) Movements of tagged snapper in the Hau-	_
raki Gulf. NZ Fish Res Div Occl Pub No 35	ł
reserves for managing Pacific coral roof fichos. Fich Pull	
91:414–427	
Eaton JW, Rawlings JB (1995) OCTAVE-A high level interac-	ł
tive language for numerical computations. Cache News	
40:11-18	
Eristhree N, Oxenford HA (2001) Home range size and use of	
Bermuda chub <i>Kyphosus sectatrix</i> (L.) in two marine	I
India Wast Indias I Fish Piel 50:120, 151	
ESRI (1999) Getting to know ArcView GIS: the geographic	Ŧ
information system (GIS) for everyone. Redlands, CA	-
Fretwell SD (1972) Theory of habitat distribution. In: Fretwell	
SD (ed) Populations in a seasonal environment. Princeton	ŝ
University Press, Princeton, NJ	
Gilbert DJ, McKenzie JR (1999) Sources of bias in biomass	
(Pagrus auratus) stock NZ Fish Assess Res Doc 99/16	c
Hilborn R. Walters C.I. (1992) Quantitative fisheries stock	
assessment: choice, dynamics and uncertainty. Chapman	
& Hall, New York	
Holland KN, Peterson JD, Lowe CG, Wetherbee BM (1993)	S
Movements, distribution and growth rates of the white	
goattish <i>Mulloides flavolineatus</i> in a fisheries conserva-	
tion zone. Bull Mar Sci 52:982–992 Holland KN, Lowo CC, Wotherbee M (1996) Meyoments and	
dispersal patterns of blue trevally (<i>Carany melampycus</i>)	7
in a fisheries conservation zone. Fish Res 25:279–292	
Kramer DL, Chapman MR (1999) Implications of fish home	7
range size and relocation for marine reserve function.	
Environ Biol Fish 55:65–79	
Lewis AR (1997) Recruitment and post-recruit immigration	
anect the local population size of coral reef fishes. Coral	1
Løkkeborg S Skajaa K Ferno A (2000) Food-search strategy	
in ling (<i>Molva molva</i> L.): crepuscular activity and use of	7
space. J Exp Mar Biol Ecol 247:195–208	
MathWorks (1998) MATLAB: the language of technical com-	
puting: computation, visualization, programming. Natick,	
MA	7
Mattnews KR (1990) A telemetric study of the home ranges	
and noming routes of copper and quillback fockfishes on shallow rocky reefs. Cap. 1 Zool 68:2243–2250	τ
	1

McGovern JC, Wyanski DM, Pashuk O, Manooch III CS,
Sederberry GR (1998) Changes in the sex ratio and size at
maturity of gag, Mycteroperca microlepis, from the
Atlantic coast of the southeastern United States during
1976–1995. Fish Bull 96:797–807

Meyer CG, Holland KN, Wetherbee BM, Lowe CG (2000) Movement patterns, habitat utilization, home range size and site fidelity of whitesaddle goatfish, *Parupeneus porphyreus*, in a marine reserve. Environ Biol Fish 59: 235–242

Parsons DM, Shears NT, Babcock RC (submitted) Radioacoustically positioned video transects: a new tool for mapping habitat change in a marine reserve. Mar Freshw Res

- Paul LJ (1967) An evaluation of tagging experiments on the New Zealand snapper, *Chrysophrys auratus* (Forster), during the period 1952 to 1963. NZ J Mar Freshw Res 1: 455–463
- Paul LJ (1976) A study on age, growth, and population structure of the snapper, *Chrysophrys auratus* (Forster), in the Hauraki Gulf, New Zealand. Fisheries Research Division, MAF. Fish Res Bull No. 13
- Roberts CM, Polunin NVC (1991) Are marine reserves effective in management of reef fisheries? Rev Fish Biol Fish 1: 65–91

Robertson DR (1988) Abundances of surgeonfishes on patchreefs in Caribbean Panama: due to settlement, or post-settlement events? Mar Biol 97:495–501

Sale PF (1978) Reef fishes and other vertebrates: a comparison of social structures. In: Reese ES, Lighter FJ (ed) Contrasts in behaviour, adaptations in the aquatic and terrestrial environments. John Wiley & Sons, New York

- Scott SG, Zeldis JR, Pankhurst NW (1993) Evidence of daily spawning in natural populations of the New Zealand snapper *Pagrus auratus* (Sparidae). Environ Biol Fish 36: 149–156
- Schoener TM (1981) An empirically based estimate of home range. Theor Pop Biol 20:281–325
- Shears NT, Babcock RC (2002) Marine reserves demonstrate top-down control of community structure on temperate reefs. Oecologia 132:131–142
- White GC, Garrot RA (1990) Analysis of wildlife radio-tracking data. Academic Press, San Diego
- Willis TJ, Babcock RC (1998) Retention and *in situ* detectability of visible implant fluorescent elastomer (VIFE) tags in *Pagrus auratus* (Sparidae). NZ J Mar Freshw Res 32: 247–254

Willis TJ, Millar RB (2001) Modified hooks reduce incidental mortality of snapper (*Pagrus auratus*) in the New Zealand commercial longline fishery. ICES J Mar Sci 58:830–841

Willis TJ, Millar RB, Babcock RC (2000) Detection of spatial variability in relative density of fishes: comparison of visual census, angling, and baited underwater video. Mar Ecol Prog Ser 198:249–260

Willis TJ, Parsons DM, Babcock RC (2001) Evidence for longterm site fidelity of snapper (*Pagrus auratus*) within a marine reserve. NZ J Mar Freshw Res 35:581–590

Willis TJ, Millar RB, Babcock RC (2003) Protection of

Moran WA (1987) Tagging confirms separate stocks of snapper in Shark Bay region. Fins 20:3–8

Munday PL, Wilson SK (1997) Comparative efficacy of clove oil and other chemicals in anaesthetization of *Pomacentrus amboinensis*, a coral reef fish. J Fish Biol 51:931–938

O'Dor RK, Andrade Y, Webber DM, Sauer WHH, Roberts MJ, Smale MJ, Voegeli FM (1998) Applications and performance of radio-acoustic positioning and telemetry (RAPT) systems. Hydrobiologia 372:1–8

14 Mar Ecol Pr	rog Ser 🔳 🗖
exploited fishes in temperate regions: high density and biomass of snapper <i>Pagrus auratus</i> (Sparidae) in New Zeoland merican comment. A park Facil 40.217	Zeller DC (1997) Home range and activity patterns of the coral trout <i>Plectropomus leopardus</i> (Serranidae). Mar Ecol
 Zealand marine reserves. J Appl Ecol 40:214–227 Woodroffe R, Ginsberg JR (1998) Edge effects and the extinction of populations inside protected areas. Science 280: 	 Prog Ser 154:65-77 Zeller DC, Russ GR (1998) Marine reserves: patterns of adult movement of the coral trout (<i>Plectropomus leopardus</i> Ser-
2126-2128	ranidae). Can J Fish Aquat Sci 55:917–924
Editorial responsibility: Otto Kinne (Editor), Oldendorf/Luhe, Germany	Submitted: January 29, 2003; Accepted: July 22, 2003 Proofs received from author(s): •••, 2003